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# On-Line Monitoring of Minor Oil Spills in Seawater Using Sediment Microbial Fuel Cells: A Preliminary Study

Jermaine Chambers<sup>1</sup>, Kaia Gomez<sup>1</sup>, Williamson Gustave<sup>1</sup>

1 University of The Bahamas

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## Abstract

In recent years microbial fuel cells have been studied for their biosensing properties. Consequently, sediment microbial fuel cells (sMFCs) have been found to be able to detect minor oil spills in freshwater. However, it was previously unknown whether sMFC properties as a biosensor would be able to produce the same results in a saltwater environment. Therefore, sMFCs at various external resistances ( $220 \Omega$ ,  $300 \Omega$ ,  $430 \Omega$ ,  $510 \Omega$ ,  $1000 \Omega$  and  $2000 \Omega$ ) were assembled to assess their ability to detect oil in seawater. The results indicated that as oil was added to the cathode of the microbial fuel cells, there was a clear and gradual decrease in the voltage output due to the oil's interference with oxygen dissolvability in seawater; However, the relationship between the change in voltage and change in time was less linear than the change observed in previous studies, and inconsistent across the different voltages. The variation observed may be due to the absence of a catalyst, such as platinum, which would have sped up the rate of the decrease in voltage. This study illustrates that sediment microbial fuel cells provide a cost effective and environmentally friendly method to detect minor oil spills in seawater.

#### Jermaine Chambers, Kaia Gomez, and Williamson Gustave\*

The School of Chemistry, Environmental, & Life Sciences, University of The Bahamas, Nassau, Bahamas

\*Correspondence: Williamson Gustave, Williamson.Gustave@ub.edu.bs

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## 1. Introduction

From the mid-1800s to the present, oil has become a valuable and treasured commodity, with offshore methods of obtaining it contributing to a large percentage of global oil production, and an even larger percentage of the oil produced is transported through or on water (Cheng & Duran, 2004). Despite precautionary measures such as epoxy coating on

underwater steel pipelines, a considerable amount of oil ends up being spilled in the ocean, causing harm to numerous ecosystems, and destroying others (Chilvers et al., 2020). While processes such as photolysis, evaporation, dispersion, and bacterial biodegradation will remove the majority of oil from the water surface, the effect on marine life is still significant, with the largest concern being for the seabird population (Kingston, 2002). Despite oil being a naturally occurring substance in the earth, volatile organic compounds (VOC) and polycyclic aromatic hydrocarbons (PAH) which are found in crude oil are toxic to humans and wildlife and can lead to the death of marine life in large quantities. Oil spills have been observed to "impede agricultural productivity and fishing", which directly impacts the economic life of populations dependent on these two industries (Osuagwu & Olaifa, 2018). This highlights the importance of developing methods to detect and mitigate the effects of oil spills, as well as taking precautionary preventative measures; however, as accidents by nature cannot be prepared for, methods of detection are paramount.

There are several remote sensing methods already being implemented to detect more prominent oil spills, such as fluorosensors (Pangilinan et al., 2016; Fuchi et al., 2022) and infrared sensors (De Kerf et al., 2020), but minor oil spills often go undetected. Minor or small oil spills generally refer to oil spills that contain a maximum volume of 6 gallons and are usually caused by leaks while refueling or discharging fuel. Despite the small amount of oil in minor spills, the toxic and flammable effects are still present, as the percentage of crude oil made up by polycyclic aromatic hydrocarbons was found to be up to 7%, and safe levels of PAHs is generally 0.2 milligrams per cubic meter (in air) (Neff et al., 2005).

The use of microbial fuel cells as a biosensor has become more popular overtime as it has been used for the testing and curing of wastewater, hydrogen production, seawater desalination, microbial electro synthesis, and detecting oil spills in freshwater sources (Do et al., 2020; Dai et al., 2021). Microbes placed at the anode metabolize the substrates they are placed in, releasing hydrogen ions and electrons. These electrons are transported from the anode to the cathode, and this passage is what creates an electrical current. The cathode acts as the electron acceptor, and these electrons then are used in the synthesis of water, utilizing oxygen obtained from the air, and the protons (H<sup>+</sup> ions) from the microbial metabolism. The current created from the anode to the cathode is passed through a resistor and a data acquisition module. This method can be used for on-line monitoring, which means that the information obtain can be analyzed and recorded from a computer that is attached to the data acquisition module.

A study done by Dai et al. (2021) on the monitoring of minor oil spills in natural waters using sediment microbial fuel cells (sMFC) proved that sMFC are capable of detecting oil spills; however, only freshwater sediment microbial fuel cells were used. It is important to also consider the biosensing properties of sMFC in seawater, as seawater has been established to contain varying concentrations of salts and other dissolved substances that differentiate it from fresh water. The majority of oil spills also occur in seawater, and their adverse effects are typically more prominent. Determining if there is any difference between the biosensing properties of sediment microbial fuel cells in seawater and freshwater would prove to be extremely beneficial as this would increase the versatility of the device, making it useable in all bodies of water.

As previously mentioned, there are numerous studies on the usage of microbial fuel cells as a biosensor, but literature on the feasibility of sMFC in detecting oil spills is extremely limited as this is a relatively new idea (Do et al., 2020). Microbial fuel cell-based biosensors have only started to be explored in recent decades and have been used for monitoring both

water and air quality (Cui et al., 2019). The cost effectiveness and relative simplicity of sediment microbial fuel cells has raised interest in their usage in developing countries where water quality testing for both drinking water and other purposes is more necessary (Chouler & Di Lorenzo, 2015). In the monitoring of water quality, studies on biochemical oxygen demand and toxicants are available due to the popularity of these parameters being tested, and the high sensitivity of microbial fuel cells to minor changes (Yang et al., 2016). Despite all of these qualifications, more studies are needed on detecting other foreign objects and particles in water and air to display the true value of microbial fuel cells as a biosensor. This emphasizes the importance of studying oil spill detection using sMFC, which is fundamentally different from the other toxins studied due to the fact that in this study the presence of the oil on the surface of the water is what causes an anoxic state in the water of the fuel cell due to oil preventing oxygen from dissolving in water.

Although numerous studies have explored sMFC as biosensors, they often overlook the influence of different external resistors on the biosensing performance of sediment microbial fuel cells. Moreover, most of the existing studies have focused on assessing sMFC as biosensors in freshwater environments. To date, there are no reports of sMFC deployment in seawater for use as biosensors. Therefore, this study aims to investigate the impact of seawater on the biosensing properties of sMFC, considering that salinity levels in seawater could potentially enhance or diminish the effectiveness of sMFC. Additionally, the study aims to explore the potential influence of resistance on the biosensing properties of sMFC.

## 2. Methods

#### 2.1. Sediment and Water Sources

The sediment (5-30 cm below sediment-water interface) samples were collected from an area 5 meters off the coastline at [25°00'56"N 77°17'56"W]. The seawater was collected from this same point. Both the water and sediment were immediately transported to the lab, and the sediment was wet sieved to remove as much large debris and rocks as possible.

### 2.2. sMFC Assembly and Operation

 Table 1. showing the resistance on each group of the

 sediment microbial fuel cells in ohms.

Sediment microbial fuel cell Group	Resistance on circuit (Ω)
A 1-3	220
B 1-3	300
C 1-3	430
D 1-3	510
E 1-3	1000
F 1-3	2000
Control group 1-3	Open Circuit

Twenty-one sMFC were constructed from the sediment and operated at different ER (2000, 1000, 510, 430, 300, and 220  $\Omega$ ) in triplicates for 50 days incubation. Of the twenty-one sMFC, 3 of them were designated as the control group, and left open circuit, meaning the anode and cathode were left disconnected. All of the sMFC were assembled according to a previously done study by Wang et al. (2015), with some minor alterations. Briefly, a 500 mL erlenmeyer flask was used to construct each sMFC. Square sections of carbon felt with geometric surface area of 25 cm<sup>2</sup> were used as anodes and cathodes. Carbon felt forms a conductive lattice structure of polyacrylonitrile fiber that supplies oxygen reduction reaction sites, and habitats for anode respiring bacteria (Gustave et al., 2018). Using 138g dry weight of sediment, ~ 2 cm depth of sediment was placed at the bottom of the sMFC container. Then, the anode was placed on the surface of the sediment layer and then buried with additional sediment to replicate anaerobic conditions, The cathode was placed above the sediment in aerobic conditions, fully submerged in seawater. Seawater (200mL) was added to flood the sediment. No external carbon source or electron acceptors were added to the sMFC during the entire operational period.

#### 2.3. Oil Spill Monitory Experiment

The cathode was fully submerged, and the fuel cells were aerated until reaching a state of quasi stabilization before they were connected to a digital multimeter (BTMETER BT- 90EPC Digital Multimeter) which was used to record the current between the anode and cathode. The multimeter was synced to a laptop via PC Link that allowed real time data collection. Each volume of motor oil (1mL, 1.5mL, 2.5mL, 5mL) was drawn from a graduated cylinder filled with oil for easy access. The oil was gently added to the fuel cell above the cathode via a micropipette to avoid interruption of the biofilm that had developed over the cathode. The fuel cells were then monitored for 60 minutes while the data was computed via the digital multimeter. After the 60-minute time period had been completed, oil was added to increase the volume for the next 60-minute time period, for total oil volumes of 1mL, 2.5mL, 5mL and 10mL. This process was repeated until the fuel cell had been monitored for 60 minutes with a volume of 10 mL of oil at the surface of the water. This process was repeated for each volume of oil for the most stable fuel cell from each group. The data collected was plotted in excel. The experiments were all done at room temperature around 1 PM to avoid variation in sediment microbial fuel cell productivity.

## 3. Results and Discussion



### 3.1. Voltage generation in the sMFC with Varied External Resistors

Figure 1. It shows the average voltage of each resistance over the course of the experiment.

The voltage increased steadily until the constant aeration ceased, after which the voltage dropped significantly. The control group exhibited a more drastic change at this time than the other fuel cells because they were open circuit, and colonies of electrophilic bacteria had not been cultivated. The average voltage of the control group decreased to a little over 300 mV, then suddenly jumped up to 700mV and stabilized there for the remainder of the study. This can be explained by the fact that the water in the fuel cells had been reduced after aerating ceased. When water that had not been reduced was added to the fuel cell to make up for the gradual water loss to evaporation, the voltage rapidly increased to a baseline of over 700mV. After the majority of the fuel cell groups had reached a semblance of equilibrium, group C still exhibited a large degree of volatility. This can be attributed to the fact that the average voltage of the group, especially during the experimental phase when oil was being added to the fuel cells. In this study each group of sediment microbial fuel cells was loaded with their own external resistor. The external resistors used were 220  $\Omega$ , 300  $\Omega$ , 430  $\Omega$ , 510  $\Omega$ , 1000  $\Omega$ , and 2000  $\Omega$ .

#### 3.2. Cell Internal Resistance and Power Density

The power density of the fuel cells was determined using the formula  $P = \frac{E_{cell}^2}{A_{an}R_{ext}}$  in accordance with the study done by Logan et al. (2006), where  $E_{cell}^2$  is the voltage of the cell squared,  $A_{an}$  is the area of the anode, and  $R_{ext}$  is the external

resistance of the cell. In this particular study, the area of the anode is equal to the area of the cathode, so these two values are interchangeable.



Figure 2. It shows the polarization curve of each cell and the corresponding power curves.

Graphs of the polarization curves (square symbols) and the power density (triangular symbols) of each fuel cell were plotted to determine the performance of each resistance group of cells (Figure 2). In this study, the power current curves were created by varying the external resistance of each fuel cell from 10K  $\Omega$  to 10  $\Omega$  using a variable resistor, and the figures obtained were plotted in figure 2 to compare the values for each group. The maximum power density for the sediment microbial fuel cells at varying resistances was 60.64, 71.97, 87.94, 97.08, 106.50, and 48.84 mW/cm<sup>3</sup> for 220  $\Omega$ , 300  $\Omega$ , 430  $\Omega$ , 510  $\Omega$ , 1000  $\Omega$  and 2000  $\Omega$  respectively. The group exhibiting the highest power density was group E, with an external resistance of 1000  $\Omega$ . This indicates that an external resistance of around 1000  $\Omega$  is optimal for sediment microbial fuel cell performance.

## 3.3. Effect of Oil Doping on Voltage Output



Figure 3. A-F, showing the change in voltage over time as oil was injected at the cathode.

Studies have shown that the addition of oil to the cathode of a sMFC can cause the voltage to decrease over time. The reason for this evident decrease in voltage as oil is added is because oil prevents oxygen from dissolving into the water (Dai et al. 2021). The microbes in the sediment are exoelectrogenic, meaning that they transfer electrons externally. Due to this inability to internally transfer electrons, the electrons are instead transferred to the anode, which is made of conductive carbon felt. The titanium wire woven through the anode then transfers these electrons to the cathode, which was originally in an aerobic environment. As oxygen is the final electron acceptor for the respiration being carried out by microbes at the anode, when oil is covering the air-water interface, the availability of dissolved oxygen (DO) is a limiting factor for the rate at which electrons can pass through the titanium wire that completes the circuit. This means that as more oxygen molecules are turned into  $H_2O$  at the cathode as a result of binding with protons (hydrogen atoms released from catabolism of substrate at the anode) and electrons that traveled to the cathode from the microbial organisms in the sediment, there are less available oxygen molecules to accept electrons. With the constant supply of oxygen being cut off

by the presence of oil, the voltage will continue to drop until it reaches 0, at which point the sMFC can be classified as dead. This is illustrated in figure 3 A-F, as the graphs all show an eventual, albeit inconsistent decrease in voltage in all the fuel cells after oil was added to the surface of the water at the cathode.

However, unlike the previous studies done on this topic, the relationship between increase in time and decrease in voltage was less linear. This could be attributed to the absence of a catalyst in this experiment, which would have greatly amplified the relationship as shown in the study conducted by Dai et al. (2021). The absence of a catalyst may also be the reason why the hypothesized relationship between the volume of oil used and the change in voltage over time was not consistent across all graphs. The issue with catalysts such as platinum, which was used in the experiment done by Dai et al. (2021) in the form of a powder coating over the cathode, is that they are typically toxic to the environment if not properly handled and disposed of. Furthermore, catalysts such as platinum are not only hard to transport but also fairly costly. This study serves as proof that there is a more cost effective and environmentally friendly way to use sediment microbial fuel cells for the detection of oil spills.

In the study done by Dai et al in 2021, Service Pro 15W-40 engine oil was used, which is significantly thicker than the Pennzoil 5W-20 that was used in this research. This also could have affected the results of the study, and future research should examine a potential relationship between the viscosity of oil and oxygen insolubility in this context. Saltwater typically has a lower concentration of dissolved oxygen than freshwater, due to the high salinity levels and subsequent increase in total dissolved solids (Al-Zubaidi et al., 2021), so in this experiment it is possible that the bacteria productivity increased drastically when the water was saturated with oxygen beyond its usual capabilities when they were aerated.

The decrease in voltage was almost immediate in all graphs; however, graphs C, D and E exhibited the most notable change over a brief period of time. This is consistent with the power density curves that indicated that a resistance of 1000  $\Omega$  (group E) is optimal for sMFC efficiency, followed by 510  $\Omega$  (group D) and 430  $\Omega$  (group C). These results can be explained by the fact that as the resistance dictates microbial activity, the group with a resistance of 2000  $\Omega$  did not deplete the DO concentration as rapidly due to the resistance limiting the rate of microbial metabolism.

The fuel cells with lower resistances such as figure 3 A and B were prone to burn out faster due to the fact that a lower resistance correlates to a high current, as shown in the formula:  $Current = \frac{Voltage}{Resistance}$ . Without a limit to the flow of electrons, the productivity of the cells would be too great, and the metabolizable substrate at the anode would be rapidly

depleted. On the other hand, if the ER was set too high, the growth of the bacterial colonies of microbes at the anode will be severely limited by the low flow of electrons due to the high resistance. In figure 3 A, the graph displays high volatility within the first hour after 1 mL of oil was added to the cathode. This can be attributed to the fact that 1 mL of oil was unable to fully cover the interface between air and water, and the amount of oxygen being dissolved varied until more oil was added, at which point it began to stabilize, then decrease rapidly as DO concentration became the limiting factor.

## 5. Conclusions

The results of this study show that as oil was inserted at the cathode of the sediment microbial fuel cell, the voltage decreased. This proves that sediment microbial fuel cells can be used for the detection of minor oil spills in seawater. As the presence of oil directly impacts the solubility of oxygen in water, there is a direct relationship, albeit non-linear, between the decrease in voltage and the increase in time after oil has been added at the cathode. Although previous studies, such as the study done by Dai et al. (2021), have suggested that the response time may be too long to be able to mitigate all the negative effects of the oil spill, this study exhibited an immediate detection of interference via oil at the airwater interface. This shows that sediment microbial fuel cells still provide a cost effective, and now environmentally safe way to detect oil spills that may have gone undetected without their utilization.

This study faced several limitations, however, in the form of time and resources. The time frame of this study did not allow for the oil doping of each replicate. Furthermore, the absence of a catalyst may have significantly altered the effects of oil doping on the sediment microbial fuel cells. When further studies are conducted on this matter, more time should be allotted to the experimental phase of the experiment, and each replicate should be used. Further research should also seek to implement a safe field study, to determine the nature of the sMFC assembly for detecting oil spills in a marine environment. Finally, it is recommended that a catalyst be obtained so that the effects of the catalyst on the oil biosensing properties of the sediment microbial fuel cell can be observed, to fully determine whether a catalyst is necessary or not. Additional research should be conducted to bring the biosensing capacities of sediment microbial fuel cells to their full potential.

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## References

- Al-Zubaidi, H. a. M., Naje, A. S., Al-Ridah, Z. A., Chabuck, A., & Ali, I. M. (2021). A statistical technique for modelling dissolved oxygen in salt lakes. *Cogent Engineering*, 8(1). <u>https://doi.org/10.1080/23311916.2021.1875533</u>
- Barron, M. G., Vivian, D. N., Heintz, R. A., & Yim, U. H. (2020). Long-Term Ecological Impacts from Oil Spills: Comparison of *Exxon Valdez*, *Hebei Spirit*, and Deepwater Horizon. *Environmental Science & Amp; Technology*, *54*(11), 6456–6467. <u>https://doi.org/10.1021/acs.est.9b05020</u>
- Cheng, L., & Duran, M. A. (2004). Logistics for world-wide crude oil transportation using discrete event simulation and optimal control. *Computers & Chemical Engineering*, 28(6–7), 897–911. <u>https://doi.org/10.1016/j.compchemeng.2003.09.025</u>
- Chilvers, B. L., Morgan, K., & White, B. (2020). Sources and reporting of oil spills and impacts on wildlife 1970–2018. Environmental Science and Pollution Research, 28(1), 754–762. <u>https://doi.org/10.1007/s11356-020-10538-0</u>
- Chouler, J., & Di Lorenzo, M. (2015). Water Quality Monitoring in Developing Countries; Can Microbial Fuel Cells be

the Answer? Biosensors, 5(3), 450-470. https://doi.org/10.3390/bios5030450

- Cui, Y., Lai, B., & Tang, X. (2019). Microbial Fuel Cell-Based Biosensors. *Biosensors*, 9(3), 92. <u>https://doi.org/10.3390/bios9030092</u>
- Dai, Z., Yu, R., Zha, X., Xu, Z., Zhu, G., & Lu, X. (2021). On-line monitoring of minor oil spills in natural waters using sediment microbial fuel cell sensors equipped with vertical floating cathodes. *Science of the Total Environment*, 782, 146549. <u>https://doi.org/10.1016/j.scitotenv.2021.146549</u>
- De Kerf, T., Gladines, J., Sels, S., & Vanlanduit, S. (2020). Oil spill detection using machine learning and infrared images. *Remote Sensing*, 12(24), 4090. <u>https://doi.org/10.3390/rs12244090</u>
- Del Campo, A. G., Cañizares, P., Lobato, J., Rodrigo, M. A., & Morales, F. S. (2014). Effects of External Resistance on Microbial Fuel Cell's Performance. *The Handbook of Environmental Chemistry*, 175–197. <u>https://doi.org/10.1007/698\_2014\_290</u>
- Do, M. H., Ngo, H. H., Guo, W., Chang, S. W., Nguyen, D. D., Liu, Y., Varjani, S., & Kumar, M. (2020). Microbial fuel cell-based biosensor for online monitoring wastewater quality: A critical review. *Science of the Total Environment*, 712, 135612. <u>https://doi.org/10.1016/j.scitotenv.2019.135612</u>
- Fuchi, T. R., Mokam, L., & Kombe, T. (2022). Drone equipment and configuration for crude oil spill detection in water.
   *European Journal of Technology*, 6(3), 1–14. <u>https://doi.org/10.47672/ejt.1127</u>
- Gustave, W., Yuan, Z., Sekar, R., Ren, Y., Chang, H. S., Liu, J., & Chen, Z. (2018). The change in biotic and abiotic soil components influenced by paddy soil microbial fuel cells loaded with various resistances. *Journal of Soils and Sediments*, *19*(1), 106–115. <u>https://doi.org/10.1007/s11368-018-2024-1</u>
- Hamilton, T. L., Bryant, D. A., & Macalady, J. L. (2015). The role of biology in planetary evolution: cyanobacterial primary production in low-oxygen Proterozoic oceans. *Environmental Microbiology*, *18*(2), 325–340. <u>https://doi.org/10.1111/1462-2920.13118</u>
- Kingston, P. F. (2002). Long-term Environmental Impact of Oil Spills. Spill Science & Amp; Technology Bulletin, 7(1–2), 53–61. <u>https://doi.org/10.1016/s1353-2561(02)00051-8</u>
- Lesser, M. P. (2004). Experimental biology of coral reef ecosystems. *Journal of Experimental Marine Biology and Ecology*, 300(1–2), 217–252. <u>https://doi.org/10.1016/j.jembe.2003.12.027</u>
- Logan, B. E., Hamelers, B., Rozendal, R. A., Schröder, U., Keller, J., Freguia, S., Aelterman, P., Verstraete, W., & Rabaey, K. (2006). Microbial Fuel Cells: Methodology and technology. *Environmental Science & Technology*, *40*(17), 5181–5192. <u>https://doi.org/10.1021/es0605016</u>
- Neff, J. M., Stout, S. A., & Gunster, D. G. (2005). Ecological Risk Assessment of Polycyclic Aromatic Hydrocarbons in Sediments: Identifying Sources and Ecological Hazard. *Integrated Environmental Assessment and Management*, 1(1), 22. <u>https://doi.org/10.1897/ieam\_2004a-016.1</u>
- Osuagwu, E. S., & Olaifa, E. (2018). Effects of oil spills on fish production in the Niger Delta.*PLOS ONE*, *13*(10), e0205114. <u>https://doi.org/10.1371/journal.pone.0205114</u>
- Pangilinan, M. N., Anacan, R. M., & Garcia, R. G. (2016). Design and Development of an Oil Spill detection and Transmission System Using Artificial Illumination Using LEDs. <u>https://doi.org/10.1109/isms.2016.61</u>
- Save The Bays: Stop the sale of Equinor until oil spill clean-up done (2023, February 23). The Tribune.

http://www.tribune242.com/news/2023/feb/23/save-bays-stop-sale-equinor-until-oil-spill-clean-/

- Wang, N., Chen, Z., Li, H., Su, J., Zhang, Y., & Zhu, Y. (2015). Bacterial community composition at anodes of microbial fuel cells for paddy soils: the effects of soil properties. *Journal of Soils and Sediments*, *15*(4), 926–936. <u>https://doi.org/10.1007/s11368-014-1056-4</u>
- Yang, W., Wei, X., & Choi, S. (2016). A Dual-Channel, Interference-Free, Bacteria-Based Biosensor for Highly Sensitive Water Quality Monitoring. *IEEE Sensors Journal*, *16*(24), 8672–8677. <u>https://doi.org/10.1109/jsen.2016.2570423</u>
- Zhou, S., Huang, S., Li, Y., Zhao, N., Li, H., Angelidaki, I., & Zhang, Y. (2018). Microbial fuel cell-based biosensor for toxic carbon monoxide monitoring. *Talanta*, *186*, 368–371. <u>https://doi.org/10.1016/j.talanta.2018.04.084</u>